



# Cell Proliferative and Cell Cycle Effects of Atrazine Using Human Breast Cell Lines

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## Abstract

Breast cancer is one of the leading causes of cancer related deaths in women. Several risk factors can increase breast cancer occurrences and strong evidence has shown that exposure to endocrine disrupting chemicals including atrazine can lead to breast cancer etiology. This study examined the effect of low level/environmentally relevant concentrations of atrazine on two breast cell lines (MCF-7, an estrogen responsive breast cancer cell line and MCF-10A, a non-cancerous human breast epithelial cell line) by studying cytotoxicity, proliferation and cell cycle events. To study cytotoxicity, cells were exposed to atrazine within a concentration range of 0.315 µg/L to 100,000 µg/L for 96 h and cell proliferation as well as the LC50 were calculated at 24, 48, 72 and 96 h using the RT-CES. Cell cycle was studied by exposing the cell to 3 µg/L atrazine, and nuclei was isolated and analyzed using a BD FACSCalibur. The percentage nuclei in the cell cycle was calculated using Modfit LT 3.0. Results showed that there was an overall decreased in LC50 over time in both cell lines. Exposure of MCF-7 and MCF-10A to 3 µg/L of atrazine for 96 h, stimulated an increase in cell proliferation compared to control. Lower concentrations of atrazine showed even stronger impact on cell proliferation. However, ATR induced more cell proliferation in MCF-7 than in MCF-10A. Significant increases were recorded in the percent nuclei in G1 and G2 phases of MCF-7 cells exposed to ATR. Increase in the percent nuclei was only recorded in the G1 phase in MCF-10A cells exposed to ATR. This adds to the body of evidence that ATR may indeed play a significant role in the onset and progression of cancer.

**Keywords:** MCF-7; MCF-10A; Atrazine; Breast Cancer

## Introduction

Breast cancer is the most frequent cancer among women, and it impacts 2.1 million women each year [1]. It is estimated that 627,000 women died from breast cancer (approximately 15% of all cancer deaths in women) in 2018 [1]. Although breast cancer occurrence is rare among men, it is estimated that 2,670 will be diagnosed with breast cancer in 2019 [2]. The etiology of breast cancer has been linked to several factors including family history/hereditary of BRCA1 and BRCA2 genes, hormone replacement therapy, obesity and alcohol [3-5]. In addition, studies have shown that ovarian hormones, including estrogen and progesterone, may increase breast cancer risk by affecting rates of cell proliferation in the breast or by supporting enhanced cellular growth induced by estrogen dependent tumors [6]. Furthermore, exposure to environmental pollutant including endocrine disruptors chemicals (EDC) have been implicated as potential risk factor in developing breast tumors [7].

Environmental pollutants including herbicides have been shown to affect cell proliferation and therefore lead to the onset of diseases [8,9]. Low level herbicide such as atrazine (ATR) exposure can

result in many adverse effects causing DNA damage, chromosomal aberration, cell cycle perturbation, as well as reproductive and developmental alteration [8,10,11]. Atrazine (2-chloro-4-ethylamino-6-isopropyl-amino-s-triazine) is a ubiquitous broad-spectrum herbicide frequently used in corn and soy fields. ATR contaminates water sources via agricultural runoffs (Solomon et al., 1996). More than thirty-four million kg of ATR are applied each year in the United States. Even though the U.S. Environmental Protection Agency (EPA) enforces a maximum contaminant level (MCL) of ATR at 3.0 µg/L (3 ppb), studies have revealed that ATR concentrations in some areas have exceeded the MCL in public drinking water sources, and the concentration in groundwater could range from below 3 µg/L to as high as 700 µg/L [12-14]. In 2003 the European Union banned the use of ATR because of its prevalence in drinking water contamination; however, the EPA permitted its continued use in the US that same year [15].

ATR is believed to be a potent endocrine disruptor [16,17] that can cause demasculinization and feminization [18] and can also induce adverse effects on different biological systems. For

example, ATR can affect the reproductive system [19,20] the central nervous system [21], and the immune system [22]. Since ATR exposure has been associated with cancer development, increased cell proliferation is an expected outcome after exposure as demonstrated in human intestinal epithelial cells [23]. In fact, a study by [24] showed that exposure to ATR induced cell proliferation in human BG-1 and 2008 ovarian cancer cells in a concentration dependent manner. Epidemiological studies have also related increased risk of ovarian cancer in female farm workers in Italy [25] and increased risk of breast cancer in the population of Kentucky to ATR exposures [26]. ATR does not bind or activate the classical estrogen receptor (ER) [24], but scientists have speculated other estrogen related pathways. For example, ATR may induce aromatase (CYP19) activity, converting testosterone and related hormones to estrogens, thus it increases estrogen levels [27]. Other studies have contradicted the endocrine disrupting effects of ATR. Studies funded by Syngenta concluded that there was no causal relationship between exposures to ATR and development of breast cancer [28]. Contradictory results were also found in another study in which higher levels of mixed pesticides, including ATR, were associated with increased breast cancer in one rural county in the UK, but not in a nearby county [29]. ATR is also known to induce alterations in normal cell cycle progression. Studies by Freeman and Rayburn [30] reported that there was a decrease in cells in the G2 phase when Chinese Hamster Ovary (CHO) cells were exposed to ATR [31]. also saw similar results when Hep2G cells were exposed to ATR.

Even though a few researchers have studied the Proliferative effects of ATR using different cell lines, studies of the effects of ATR on cell cycle in human breast cell lines are scarce. Since endocrine disruptors like ATR act like natural hormones which are available in low doses, this study explored the effect of low level/ environmentally relevant concentrations of ATR on two human breast cell lines (MCF-7, an estrogen responsive breast cancer cell line and MCF-10A, a non-cancerous human breast epithelial cell line) by studying cytotoxicity and cell cycle events. The study will help scientists understand whether ATR will increase the risk of women who have already been diagnosed with breast cancer and those who have not.

## Material and Methods

### Reagents

MCF-7 and MCF-10A cell lines were purchased from American Type Culture Collection (ATCC, Manassas, VA, USA). Minimum Essential Medium (MEM) alpha 1x, Dulbecco's Phosphate Buffered Saline (PBS), Dulbecco's Modified Eagle's Medium (DMEM), MEM without phenol, and penicillin streptomycin were purchased from GIBCO Invitrogen (Grand Island, NY, USA). Trypsin-EDTA and fetal bovine serum (FBS) were purchased from ATCC. All other supplements were purchased from Cambrex Bio Science, Walkersville, Inc., USA.

### Cell Culture

Both MCF-7 and MCF-10A cell lines were incubated at 37°C in 5% CO<sub>2</sub> and humidified atmosphere. When MCF-7 cells reached 75-80 % confluency, they were washed with phosphate buffer saline (PBS), trypsinized with 3 ml of 0.25% w/v trypsin, 0.53 mM EDTA and incubated at 37°C for 5 minutes. For MCF-10A cells, 3.0 ml of a 0.05% trypsin, 0.53 mM EDTA solution was added and cells were incubated at 37°C for 15 minutes. At the end of the incubation period, 3 ml of a solution of 0.1% soybean trypsin inhibitor was added to the MCF10 A to neutralize the trypsin. Both cell types were centrifuged at 3000 rpm for 2 minutes at 4°C. Both cell lines were diluted with MEM, washed with saline and centrifuged at 5000 rpm for 5 minutes at 4°C. Cells were counted with Beckman counter and used for exposure assays.

### Description of Exposure Studies

**Real-Time Cell Electronic Sensing (RT-CES):** Cell viability and cytotoxicity effects of ATR on MCF-7 and MCF-10A cells were determined by RT-CES assay (ACEA Biosciences Inc., San Diego, CA, USA). Cell viability was monitored every 10 min for 96 h by the detection of cell impedance as a measure of cell number morphology and adherence. Continuous recording of impedance in cells was reflected by cell index (CI) value which corresponds to cell growth [31].

**Cytotoxicity of ATR on MCF-7 and MCF-10A cell lines Using RT-CES:** To determine the cytotoxicity of the ATR, MCF-7 and MCF-10A cells were seeded in a 16x E-plate device and grown in the incubator. After 24 h, the cells were treated with ATR. Serial dilution as described by [9] were followed. To create a negative control, the last row of cell culture plate contained the media and cells but was not exposed to ATR. Cells were exposed to ATR within a concentration range of 0.315 µg/L to 100,000µg/L for 96 h and the lethal concentration at which 50% of the cells died (LC50) was calculated at 24, 48,72 and 96 h. In addition, CI was derived to represent cell status based on measured electrical impedance. A rise in CI (rise in electrical impedance) indicated cell proliferation and a drop in CI (drop in electrical impedance) indicated cell death. The experiment was conducted three separate times. ANOVA followed by mean separation using LSD at  $\alpha = 0.05$  was also calculated [32].

### Cell Cycle Analysis

Both cell lines were exposed to 3 µg/L environmentally realistic concentrations of ATR. Eighty pico-gram/ml of estrogen was used as a positive control for measuring estrogenic activity of the cell lines. After 96 h exposure period, nuclei were isolated from exposed and control cells then stained with propidium iodide (PI) using a hypotonic lysis solution [33]. Briefly, cells were washed with sterile 1% PBS, followed by the addition of 1.5 ml of the PI solution (0.05mg/ml, 0.1 Triton X 100, 0.1 % sodium citrate, 7 unites /ml RNase). Plates were refrigerated and tilted every 3-5 minutes to release the nuclei from the lysed cells. The samples were

filtered through a 53-um mesh filter and kept on ice until analysis. The refrigeration time differed for each cell line. MCF-7 cells were refrigerated for one hour and MCF-10A cells were refrigerated for 6 hours as nuclei from MCF10A takes longer time to be collected than does MCF-7 cells. The isolated nuclei were then analyzed using a BD FACSCalibur equipped with 4-Color filters; 530 nm (FITC), 585 nm (PE/PI), 670nm (PerCP) and 661 nm (APC), (San Diego, LA, USA). The excitation wavelength (488 nm) was provided by a 5 W argon ion laser. Approximately 20,000 nuclei per sample were analyzed. The percentage of nuclei of the cell cycle was calculated using Modfit LT 3.0 (Verity Software House Inc., Topsham, ME, USA). ANOVA followed by mean separation using LSD at  $\alpha = 0.05$  was performed on the percentage of nuclei in the G1, S (synthesis), and G2 phases of the cell cycle.

**Results**

**Cytotoxicity of ATR on MCF-7 and MCF-10A cell lines**

Median lethal concentration (LC50) is the most widely used criterion for acute toxicity testing. LC50 is described as the concentration of a substance that kills 50% of the test organisms after a specific length of exposure, usually 96 h. The LC50 for MCF-7 cells exposed to ATR were 14,960, 22,100, 14,700 and 14,100  $\mu\text{g/L}$  for 24 h, 48 h, 72 h and 96 h respectively. The LC50 for MCF- 10A

were 36,660, 9,870, 9,940 and 1,200  $\mu\text{g/L}$  for 24 h, 48 h, 72 h and 96 h respectively (Table 1).

**Table 1:** LC50 ( $\mu\text{g/L}$ ) of ATR on MCF-7 and MCF-10A cell lines.

| Mean $\pm$ Std. Dev |                    |                   |
|---------------------|--------------------|-------------------|
| Time                | MCF-7              | MCF10-A           |
| 24                  | 14960 $\pm$ 0.424B | 36660 $\pm$ 0.23C |
| 48                  | 22100 $\pm$ 0.694A | 9870 $\pm$ 0.12B  |
| 72                  | 14700 $\pm$ 0.249B | 9940 $\pm$ 0.09B  |
| 96                  | 14100 $\pm$ 0.282B | 1200 $\pm$ 0.16A  |

\*Means within a column lacking a common letter are significantly difference ( $p \leq 0.001$ )

**Cell Proliferation After ATR Exposure**

Cell lines were exposed to ATR concentrations ranging from 0.315  $\mu\text{g/L}$  to 100,000 $\mu\text{g/L}$ . Results showed that exposure of MCF-7 breast cancer cells to 3 $\mu\text{g/L}$  ATR (EPA MCL) for 96 h stimulated an increase of 20% cell proliferation compared to control (Figure 1). Lower concentrations than the EPA MCL for ATR showed even stronger impact on cell proliferation, for example, 1.5  $\mu\text{g/L}$  increased cell proliferation by 31% as compared to control. ATR concentration of 0.75  $\mu\text{g/L}$  also caused growth to increase by 11%. ATR concentrations higher than 25,000  $\mu\text{g/L}$  induced cell death in MCF-7.

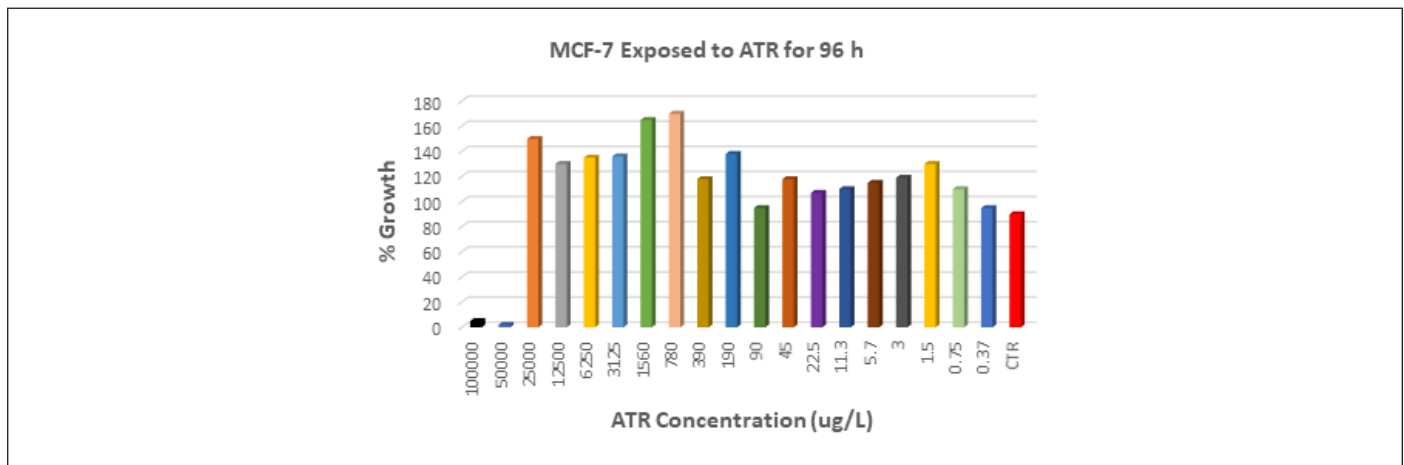


Figure 1: Breast Cancer Cell Line (MCF-7) Exposed to ATR for 96 h.

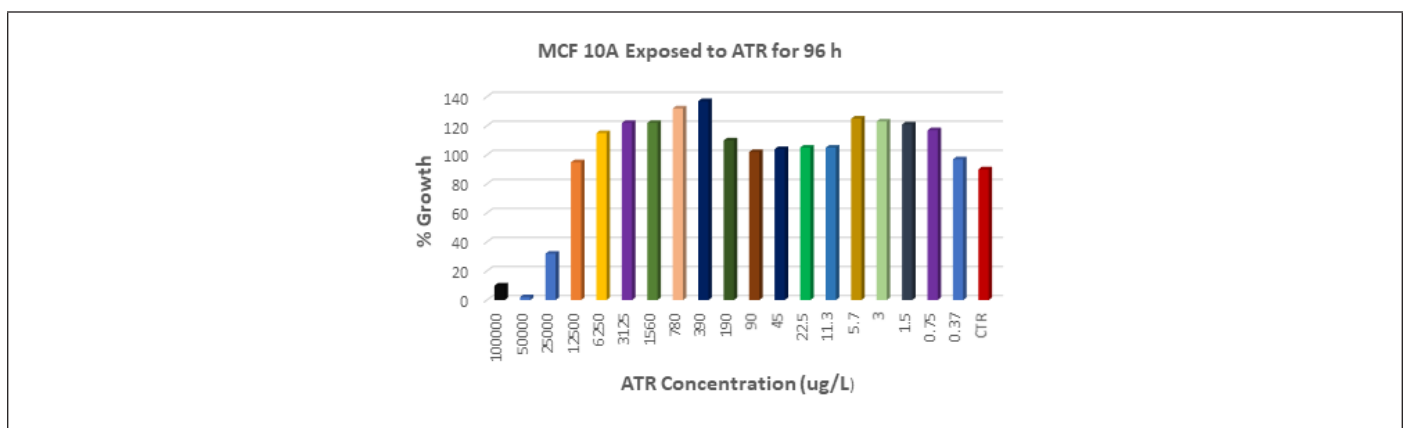


Figure 2: Normal Breast Cell Line (MCF-10A) Exposed to ATR for 96 h.

After 96 h of exposure, ATR concentrations of 3 µg/L stimulated 25% increase in cell proliferation in MCF-10A breast cells compared to control (Figure 2). Lower concentrations as 1.5 µg/L and 0.75 µg/L showed almost same stimulatory effect. Higher concentrations of ATR: 390 µg/L and 780 µg/L induced the highest increase in cell growth; 37% and 32% more than control respectively. Concentrations higher than 12500 µg/L induced cell death.

### Impact of ATR on cell cycle

To determine the effect of ATR on cell cycle, both Cell lines were exposed to the EPA MCL concentration of ATR for 96 h. The result indicated that there was a significant increase in the % nuclei in the G1 and G2 phases of MCF-7 cell lines; however, there was no increase in the S phase. MCF-10 A cells showed a significant increase in the % nuclei in the G1 phase but there was not increase in % nuclei in the G2 or S Phase (Table 2, Table 3 & Table 4).

**Table 2:** G1 Phase Analysis of both cell lines after 96 h of exposure.

| TRT | *% G1 phase ± Std. Dev |                |
|-----|------------------------|----------------|
|     | MCF-7                  | MCF10-A        |
| CTR | 94.57 ± 0.92 C         | 75.88 ± 1.75EF |
| EST | 94.59 ± 1.11 C         | 75.89 ± 0.70F  |
| ATR | 98.21 ± 0.41 A         | 78.7 ± 0.62D   |

\*Means within a column lacking a common letter are significantly difference ( $p \leq 0.001$ )

**Table 3:** S Phase Analysis of both cell lines after 96 h of exposure.

| TRT | * % S phase ± Std. Dev |                |
|-----|------------------------|----------------|
|     | MCF-7                  | MCF10-A        |
| CTR | 4.58 ± 0.78 C          | 22.26 ± 0.11A  |
| EST | 5.02 ± 1.06 BC         | 21.79 ± 0.31AB |
| ATR | 0.09 ± 0.05 E          | 20.97 ± 0.47 B |

\*Means within a column lacking a common letter are significantly difference ( $p \leq 0.001$ )

**Table 4:** G2 Phase Analysis of both cell lines after 96 h of exposure.

| TRT | * % G2 phase ± Std. Dev |                |
|-----|-------------------------|----------------|
|     | MCF-7                   | MCF10-A        |
| CTR | 0.85 ± 0.14B            | 1.86 ± 1.86 CD |
| EST | 0.39 ± 0.25 DE          | 2.32 ± 0.63 BC |
| ATR | 1.70 ± 0.41 A           | 0.33 ± 0.33 DE |

\*Means within a column lacking a common letter are significantly difference ( $p \leq 0.001$ )

### Discussion

Cytotoxic effects of environmentally relevant concentrations of ATR on breast cancer cells (MCF-7) and normal breast cells (MCF-10A) were assessed using RT-CES. The E-plates in the instrument are equipped with microelectrodes which causes changes in electrical impedance. Thus, higher cell index indicate that more cells are bound to the microelectrodes. ATR action was recorded online after cell attachment and continuous information about

growth, morphological changes and cell death were collected in real time.

LC50 is the concentration of a substance that kills 50% of its test subjects when administered in a single dose. Acute toxicity data are very important because they are used to set guidelines for regulatory measures (PHAGM, 2005). The current study showed that the effect of ATR was time dependent; in general, toxicity increased as exposure time increased. Although, we found a significant decrease in ATR toxicity on MCF 7 after 48 h. However, a significant increase of ATR on MCF 7 cells was recorded after 72 h. This is in accordance with the findings that showed that the toxicity of ATR in *C. Puctatus* is both time and concentration dependent [34] Although, their study found the LC50 of a 96 h ATR exposure for *C. puctatus* to be 42,381 µg/L which was quite high compared to our findings. The LC50 found in this current study (14, 100 and 1,200 µg/L for MCF-7 and MCF-10A respectively) is closer to results obtained from previous studies [35-37] that reported the LC50 for ATR after 96 h exposure to be 16,000, 18800 and 9370 µg/L for *Lepomis macrochirus*, *Cyprinus carpio* and *Oreochromis niloticus*, respectively. Different cytotoxicity studies have shown that ATR exposure caused changes in erythrocytes membranes, DNA damage, mitochondrial dysfunction, cell autophagy and apoptosis [38,39]. Ultimately, the cellular changes may lead to severe adverse health effects including cancer.

It has since been purported that ATR, a potent EDC, could induce proliferative effect on breast cancer cells through aromatase activation [27] as well as GPR30 binding [40]. The current study indicates that both cell lines exposed to low environmentally relevant concentrations of 3 ATR for 96 h stimulated cell proliferation. However, exposure of cell lines to the concentration of ATR higher that 25,000 µg/L (MCF-7) and 12,500 µg/L (MCF-10A) caused cell death. It is important to note however, that the proliferative effect of ATR in the current study was greater in MCF-7 than in MCF-10A. The increased proliferation may be attributed to the fact that MCF-7 is estrogen responsive and it yielded more response to ATR's endocrine disrupting effects. Our findings are in line with the study that showed that ATR enhanced transcription of the aromatase gene as well as estrogen production in estrogen responsive breast cell [41,42]. Similarly, studies by [43] found that there was an increase in cell proliferation when MCF-7 cell were exposed to environmentally relevant concentrations. However, they did not record any increase in cell proliferation in MCF 10A. Even though the EPA has a set MCL of 3 ppb, it is very worrisome to know that lower levels can induce cell growth and therefore cancer cell proliferation. ATR has been classified as non-carcinogenic by the EPA, however, strong evidence suggests otherwise. Scientist continue to record more environmental and human adverse effects due to exposure [43] including increased risk of ovarian cancer in female farm workers in Italy [25] and increased risk of breast cancer in the population of Kentucky to ATR exposures [26].



To our knowledge, our team is the first to investigate the cell cycle effects of ATR on MCF-7 and MCF 10A cells. The cell cycle investigation in the present study buttresses the body of evidence that implicates ATR in the etiology of breast cancer as cell proliferation was confirmed by a significant increase in the % nuclei in the G1, and G2 phases in MCF-7. However, significant increase in % nuclei was only recorded in G1 phase for MCF-10A. Increase of nuclei in cell cycle confirms cell duplication and proliferation. This finding is in agreement with a study conducted in HepG2 cell line, where 100, 300, and 500 ppb ATR exposure for 48 h caused nuclei to accumulate in S phase compared to control [31]. Similarly, studies using CHO cells exposed to 43ppm ATR, showed a significant accumulation of nuclei in S phase [30]. Although an increase in the % nuclei in the G phase was recorded in the current study, unlike previous studies, we did not record any increase in the % nuclei in cells in the S phase. This variation may be attributed to the difference in the type of cell lines used in the study. Interestingly, estrogen did not elicit an increase the % nuclei in any of the cell cycle phases in both cell lines. This suggests that estrogen may not exerts its endocrine disrupting effects through cell cycle activities. It is important to also note that even though ATR and estrogen elicit endocrine disrupting effects, ATR does not bind to or activate the classical estrogen receptor, but it up-regulates the aromatase activity in estrogen-sensitive tumor cells [24].

The current study examined the effects of ATR exposure on human breast cell. Specifically, ATR toxicity. Cell proliferation activity and cell cycle events were studied by exposing cells to different concentrations of ATR including very low concentrations that can be present in drinking water. In addition, two human breast cell lines were studied rather than non-human cell lines that are frequently published. The results obtained from studying human breast cell lines may prove more relevant in understanding the etiology of breast cancer in humans. It is imperative that more studies that elucidates the possible mechanistic pathways by which ATR induces cancer in human cells- specifically breast cells, be conducted.

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